ARTEMESIA LONGINARIS has been recently exploited by commercial and artisanal fisheries in Southern Brazil. To investigate the reproductive biology of this species, monthly samples were analyzed between February/2002 and January/2003, in depths between 5 and 20 meters in surrounding area of the Barra do Rio Grande. Stock-recruitment relationships (SRR) were estimated, as well as a preliminary predictive model for recruit biomass. Reproductive effort is concentrated in spring; however, recruitment was observed year round. Reproduction in autumn is more variable, and susceptible to intense rainfall and subsequent low salinity observed in the studied area. A reproductive migration to depths farther than 15 meters was observed, and recruitment in shallower waters starts approximately one month after hatching. Mean length at first maturity was estimated in 16.76 mm and is achieved in approximately 4 months. Significant fit of stock-recruitment relationships (SRR) was observed for three models tested under adverse environmental conditions, suggesting that the reduction of spawning biomass combined with unfavorable environmental conditions may lead to recruitment overfishing. A preliminary predictive model was developed and the main factor influencing recruitment of the species is salinity in spawning season. The main factor influencing recruitment success under adverse environmental conditions is the biomass of spawning stock, while under favorable conditions salinity in spawning season plays a significant role.

PALAVRAS CHAVE: Avaliação de estoque, peneídeos, biologia pesqueira, reprodução, manejo pesqueiro.

KEYWORDS: stock assessment, penaeid shrimps, fishery biology, reproduction, fishery management

INTRODUCTION

Artemesia longinaris Bate, 1888 is a penaeid shrimp commercially exploited along its entire distribution area, from Rio de Janeiro (21°37'S), Brazil to Puerto Rawson (43°00'S), Argentina. The stiletto shrimp is usually used as food for human consumption as well as for bait in amateur fishery. This species is caught along with Pleoticus muelleri, the larger red shrimp, which is also a valuable fishery resource in Southern Brazil and Argentina (Boschi 1969; D’Incao et al. 2002). Commercial landing in Brazilian coast dates back from 1978, however, landing statistics were only available after 1982, showing remarkable increases in recent decades (Univali 2011). The growing interest and increasing fishing effort over A. longinaris, especially after the collapse of pink shrimp fishery (D’Incao et al. 2002), raise the need for new investigations about its population dynamics.

Although penaeid stocks throughout the tropics and subtropics support highly profitable fisheries (FAO 2007), little is known about the population dynamics of several species (Cha et al. 2002). Regarding population parameters, one of the most important is the size or age at first maturity, which is widely used as a biological reference point for managing exploited stocks (Gulland & Rothschild 1981, Cha et al. 2002; López-Martinez et al. 2005, Frisch 2007) since preservation of individuals smaller than mean size at first maturity increases the chance of success of the next offspring (Garcia & Le Reste 1981, King 1997).

Few previous investigations concerned on the reproduction and ovary maturation of A. longinaris.
from Southern Brazil (Nascimento 1981, Dumont & D’Incao 2004, Castillo et al. 2007). Using variation of size at first maturity, a hypothesis of two different populations was stated for this species (Nascimento 1981). However, the trait used to define morphological maturity was the presence of a spermatophore, which is presently known to be invalid as an indicator of ovarian maturity for this species, since females can carry spermatophores without being fully mature (Scelzo 1991). Dumont & D’Incao (2004) established a chromatic scale, based on ovarian microscopic traits, to identify stages of gonadal development of A. longinaris more accurately, providing a routine method to classify the ovaries. Recently, Castillo et al. (2007) investigated trends in size at first maturity in populations inhabiting distribution limits of the species.

Furthermore, investigations on penaeid shrimps biology suggest conspicuous trends in reproductive dynamics for this group, highly dependent on environmental conditions (Staples et al. 1985, Vance et al. 1998, Crocos et al. 2001). Recruitment of juveniles and larval abundance of A. longinaris has been previously assessed and marked oscillations in time and intensity of reproduction were reported, even though the main breeding season in Southern Brazil is accepted as taking place in spring (Nascimento 1981, Ruffino & Castello 1992, Calazans 2002). Previous studies suggest that main reproductive season in Argentina takes place in summer and that endogenous and environmental factors may affect reproduction of this species in that area (Boschi 1969, Christiansen & Scelzo 1971, Scelzo 1991, Petriella & Bridi 1992). In Southeastern Brazil reproduction was reported to take place in summer, associated with cold water upwelling (Costa et al. 2005).

Prediction of recruitment is also of special interest for penaeid fisheries management (Pauly 1992). Forecasting recruit abundances in penaeid populations has been considered a hard task due to their high fecundity and susceptibility to environmental changes, such as those caused by rainfall and wind stress (Garcia 1981; Gulland & Rothschild 1981, Crocos et al. 2001). Despite that, stock-recruitment relationships (SRR) have been applied to estimate recruitment of penaeid with considerable level of success (Garcia 1981, Kirkwood 1981, Pauly 1992).

There is evidence that level of recruitment is at least partly governed by environmental conditions and interactions between spawning stock and recruitment, and environmental conditions must be investigated to properly manage and predict future abundances (Pauly 1992). There are, therefore, many aspects of interaction between shrimp biology and environment that must be carefully considered, especially under low spawning biomass levels (Garcia 1981, Kirkwood 1981, Gulland & Rothschild 1981, Dall et al. 1990, Pauly 1992).

The aim of this paper is to investigate the mean size at first maturity, reproductive dynamics, stock-recruitment relationships as well as to develop a preliminary model to predict the abundance of A. longinaris recruits in the commercial fishery of extreme Southern Brazil.

MATERIAL AND METHODS

Study area

Sampling took place in surrounding area of the Barra de Rio Grande, which links the estuary to marine coastal waters through a four kilometers channel (Figure 1). This area is strongly influenced by freshwater discharge from Patos Lagoon Estuary (Ciotti et al. 1995) and discharge values of 30,300m³s⁻¹ have been reported (Garcia 1996). The estuary is located at coastal plain of Rio Grande do Sul State, Brazil (32°S, 49°W) and it is the largest choked lagoon in the world, accounting for an area of 10,360km². Drainage basin covers 201,626 km² (Asmus 1996) and rainfall as well as wind stress regulate the water movements in study area (Garcia 1996, Piola et al. 2004, 2005). Seasonal freshwater discharge presents great variability (30,300 - 47m³s⁻¹) and during El Niño Southern Oscillation (ENSO) events runoff is usually higher than mean values estimated (3,000m³s⁻¹) (Garcia 1996).
FIGURE 1 – A- Southwestern Atlantic Ocean and distribution of *A. longinaris* in coastal waters (black line). B- Closer view of Southern Brazilian coast, highlighting the surrounding area of Patos Lagoon Estuary and nine sampling stations (white dots) positioned at depths varying from 5 to 20 meters.

Data sources

Standardized data used in this paper were obtained during oceanographic cruises (L.Oc. Larus) from 1982 to 2003, in depths varying from 5 to 20 meters. To investigate reproductive pattern only data obtained during 2002/2003 were analyzed, since during previous research cruises, ovary classification had not been performed. To estimate stock-recruitment relationship (SRR), data between 1982 and 2003 were selected based on availability of information from spawning biomass and subsequent recruitment and only nine, out of twenty one years, were used. Discard of twelve years was due to gaps in collections during spawning or recruitment periods. Spawning season was considered as spring (October, November, December) and recruitment as summer (January, February, March). It is important to stress that the main reproductive event takes place in spring and commercial trawling for *A. longinaris* concentrates maximum fishing effort during summer (Univali 2011). Therefore, commercial fishery depends on recruitment originated from cohorts hatched during previous spring spawning season.

Reproductive dynamics and size at first maturity

To establish relative abundance index (CPUE) the catch unit adopted was number of individuals and the effort unit used was five minutes trawling. Mean relative abundances were statistically compared by One-Way ANOVA and subsequent *post-hoc* Tukey test (3 or more means compared) or *t* test (2 means compared), with confidence level of 95%. Assumptions of normality (Kolmogorov-Smirnov) and
homogeneity of variances (Levene) were tested prior to perform ANOVA.

Carapace length (CL) was used to describe the size structure of population, as measured from post-orbital angle to the end of mid-dorsal carapace. Size-class interval used was 0.5mm (CL). Ovary development was classified based on color and shape traits, previously established by using histological sections. Three ovarian development stages were adopted (I-im immature, II-developing, III-ripe) (Dumont & D’Incao 2004).

Recruitment patterns were obtained using the ELEFAN II routine (Pauly 1987), included in FISAT program (Gayanilo et al. 1995). Growth parameters used to estimate peaks of recruitment were obtained from von Bertalanffy growth model (VBGM). Monthly relative frequency of ripe (stage III) females was compared to the recruitment pattern generated by ELEFAN II to validate the pattern obtained. Modal groups of females were tentatively linked to visualize modal progression along the sampling period. To estimate growth parameters all the chosen cohorts were fitted to von Bertalanffy growth model (VBGM) (1938) by an automated least squares fitting procedure. The von Bertalanffy equation is given by:

\[ CL_t = CL_\infty \left[ 1 - e^{-k(t-t_0)} \right], \]

where \( CL_t \) is length at the time \( t \), \( CL_\infty \) is asymptotic carapace length, \( k \) the coefficient of growth and to the theoretical age at zero length. Longevity \( (t_{max}) \) was estimated by inverted von Bertalanffy (1938) equation, considering maximum longevity \( (t_{max}) \) as reached at 99% of the asymptotic length (D’Incao & Fonseca 2000).

Mean size at first maturity (LM) was considered as the size class interval (1mm interval) in which frequency of ripe females is 50% (King 1997). Frequency of ripe females was fitted to the logistic model by an automated least square procedure, in such a way that:

\[ P=1/(1+e^{-r(CL_1-LM)}) \]

where \( P \) is the percentage of ripe females in a given length class, \( r \) is the logistic curve slope, \( CL_1 \) is the upper limit of carapace size interval and \( LM \) is the mean length at first maturity.

Stock-recruitment relationships (SRR) and multiple regression analysis

Standardized data obtained from oceanographic cruises (L.Oc. Larus) were used to estimate spawning (females larger than LM caught during spring) and recruitment (females smaller than LM caught during summer) abundance index (g/30 minutes of trawling).

According to Gulland & Rothschild (1981) a single spawning-recruitment (SRR) curve cannot entirely describe the changes in stock biomass of an exploited shrimp population. Besides spawning biomass, environmental factors play an important role in recruitment success and therefore must be considered in stock assessment (Garcia 1981, Kirkwood 1981, Gulland & Rothschild 1981, Dall et al. 1990). Thus, two different SRR were visually established and their respective environmental parameters investigated, in such a way that one curve was estimated for favorable environmental conditions (1988, 1989, 2000, 2002) and another for adverse years, where the same spawning stock resulted in a smaller recruitment biomass (1982, 1983, 1984, 1996, 2001).

Stock-recruitment relationships (SRR) were obtained by using three different models. The Beverton & Holt (1957 apud King 1997) model assumes that recruitment achieves an asymptotic value of biomass at high spawning stock abundances and the equation that describes it is given by:

\[ R=S/(a+bS) \]

where \( R \) is the abundance of recruits in the next year, \( S \) is the spawning biomass and \( a \) and \( b \) are the parameters of the model.

The Ricker (1975 apud King 1997) equation describes a stock-recruitment relationship where recruitment achieves a maximum and decreases afterwards at high spawning stock abundances. The equation is given by:

\[ R=aS \exp(-bS) \]

where \( R \) is the abundance of recruits in the next year, \( S \) is the spawning biomass and \( a \) and \( b \) are the parameters of the model.

The last model fitted was created by Shepherd (1982 apud King 1997).

\[ R=aS/(1+(S/K)^b) \]

The parameters \( R \) and \( S \) have the same meaning as in other equations. The parameter \( a \) is the
initial slope at the origin and reflects the potential stock-recruitment relationship without the density-dependent effects. The parameter $K$ represents the threshold spawning biomass above which density-dependent effects dominate density-independent effects. Specifically, $K$ is the stock size at which recruitment is reduced to one half the level it would have been under density-independent process only (King 1997). Spawning stock and recruitment abundances were fitted to SRR models by an automated least square procedure. Mean salinity and rainfall during spring were grouped according to good and bad recruitment years (visually determined according to SRR) and then compared by a $t$ test at 5% significance level.

Multiple regression analysis was performed to elucidate main environmental and biological factors affecting the reproductive success of $A.\ longinaris$ in surrounding area of Patos Lagoon estuary. Predictors used were: total rainfall in spawning (RSS) (spring) and recruitment seasons (RRC) (summer), salinity in spawning season (SALS) (spring), salinity in recruitment (SALR) and spawning biomass (SS). Recruitment abundance in subsequent summer (RC) was considered as the dependent variable. A forward stepwise method was chosen, in such a way that most important predictors ($p<0.1$) were progressively included in regression equation. Salinity data was measured with a termosalinometer after each fishing station. Daily rainfall (mm$^3$) data was obtained from Estação Agrometeorológica da Universidade Federal de Pelotas, and seasonally grouped as previously mentioned.

RESULTS

Environmental data

Rainfall data collected for the city of Rio Grande was chosen due to higher correlation with salinity in surrounding area of Patos Lagoon ($r=0.73, b=0.049, p<0.10$). Conversely, polled rainfall data from all drainage basin did not show a close relationship with salinity. Lack of correlation may be due to complex interactions between rainfall and wind stress (Garcia 1996), resulting in a larger gap between rainfall and freshwater runoff.

Mean annual rainfall, between 1913 and 2002, in the city of Rio Grande was 1215mm$^3$. Total rainfall recorded in 2002 was the highest since 1915, reaching the value of 1915mm$^3$. Values recorded in 2002 were consistently above typical values for the area, except in July, when slightly lower rainfall was observed. Four main peaks of rainfall were observed in January (241.7mm$^3$), March (224.2mm$^3$), October (195.8mm$^3$) and December (272.7 mm$^3$) (Figure 2).

![FIGURE 2 – Trends in rainfall (mm$^3$) observed during 2002 for the city of Rio Grande (Southern Brazil), compared to average values from 1913 to 2002.](image)
Higher values of salinity were observed in summer (February= 30.2±0.39), despite of a striking reduction of values in March (24.4±1.8), as a consequence of intense rainfall during this month. Salinity tended to decrease towards winter when lowest values were recorded (September= 19.07±5.37). After the winter, salinity increased again towards December (23.1±5.83), when another reduction, associated to high rainfall, was noticed (Figures 2, 3).

As expected, temperatures were higher in summer (January= 25.64±0.78) and lower in winter (August= 13.7±0.6). From September (15.75±0.5) onwards, temperatures increased and stabilized between November (21.4±0.3) and December (21.4±0.5) (Figure 4).
Seasonal reproductive dynamics and size at first maturity

A total of 2623 females was analyzed between January/2002 and January/2003 for reproductive pattern and size at first maturity estimates. Ripe, vitellogenic and adult females were recorded year round, but mainly from July to October. Conversely, very low frequencies of ripe females were recorded in May and January/2002 (Figure 5).

Three peaks of adult females and proportion of ripe ovaries were observed along the year (Figures 5, 6). The highest density of adult females was recorded in October (38.86%) (Figure 5). In May, another high mean relative abundance of adult females was recorded (12 shrimps/min) (Figure 6). However, a very low proportion of ripe ovaries during this month was observed (<5%) (Figure 5). In January, the smallest abundance of *A. longinaris* females was observed (1 shrimp/min), also with a low percentage of mature ovaries (<5%) (Figures 5, 6).
FIGURE 6 – Monthly relative abundance (CPUE in numbers/minute) of *A. longinaris* adult females in Southern Brazil. Vertical bars represent standard deviation of means.

Trends in relative abundance of recruits coincide with pattern estimated by ELEFAN II routine, which resulted in three peaks along the year (Figure 6). Recruitment peaks presented a one-month delay after maximum values of adult females and proportion of ripe females (Figures 5, 6, 7, 8). Main recruitment event takes place in November, as indicated by high percentage (Figure 7) and relative abundance (Figure 8) of individuals smaller than LM. The other two recruitment events are less important in magnitude and took place in February and June (Figures 7, 8).

FIGURE 7 – Monthly recruitment pattern estimated by ELEFAN II routine for *A. longinaris* females in Southern Brazil. Values above grey bars indicate frequency of recruits in samples.
Smallest ripe female measured was 13.3mm and the largest 30.49mm. Estimated mean length at first maturity (LM) was 16.76mm (Figure 9, Table I) and based on length-age table, obtained from VBGM, the LM estimated corresponded to the age of approximately 4 months. The length at which 100% of females were mature was 25.0 mm, corresponding to the age of 8 months (Figure 9).

FIGURE 8 – Monthly relative abundance (CPUE in numbers/minute) of *A. longinaris* recruits (females) in Southern Brazil. Vertical bars represent standard deviation of means.

FIGURE 9 – Logistic curve of ovary maturation for estimated *A. longinaris*. Carapace length (CL mm) that corresponds to frequency of 50% is 16.76mm. Length at which 100% of females presented ripe ovaries is 25mm. External lines represent the confidence interval of estimate (95%) and internal bands the prediction interval (95%).
TABLE I – Summary of estimates obtained from least square procedure that fitted the frequency of *A. longinaris* ripe females to a logistic model, containing slope (r), length at first maturity (LM), t-values, confidence limits (CI95%) and p-values.

<table>
<thead>
<tr>
<th>Parameter</th>
<th>value</th>
<th>CI (95%)</th>
<th>p</th>
</tr>
</thead>
<tbody>
<tr>
<td>r</td>
<td>0.55</td>
<td>0.43</td>
<td>0.66</td>
</tr>
<tr>
<td>LM</td>
<td>16.76</td>
<td>16.33</td>
<td>17.19</td>
</tr>
</tbody>
</table>

Size frequency and VBGM analysis

Carapace lengths ranged from 3.1mm to 30.49mm and larger females were observed from winter to spring (July, August, September and October). During March (10.35mm), May (11.70mm), November (10.72mm) and December (10.81mm) the lowest mean CL values were recorded (Figure 10).

![Figure 10](image-url)
Analysis of CL data showed that differential length distribution of females occurred in studied area. Statistical comparison of mean length between isobaths (5-15m and 16-20m) suggests that significantly larger females (p=0.0009) inhabited deeper grounds, in isobaths greater than 15 meters (CL= 13.9 ± 3.94 mm). Conversely, smaller females (CL= 9.89 ± 3.71 mm) presented a preference for shallower waters near the beach (Table II).

### TABLE II – Trends in length (CLmm) composition grouped in two categories, from 5-15 and 16 to 20 meters.

<table>
<thead>
<tr>
<th>Depth</th>
<th>n</th>
<th>CLmm</th>
<th>Std. Dev.</th>
<th>CI (95%)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Overall</td>
<td>1006</td>
<td>12.61</td>
<td>4.24</td>
<td>12.35 12.87</td>
</tr>
<tr>
<td>5-15</td>
<td>328</td>
<td>9.89</td>
<td>3.95</td>
<td>9.46 10.32</td>
</tr>
<tr>
<td>16-20</td>
<td>678</td>
<td>13.93</td>
<td>3.72</td>
<td>13.65 14.21</td>
</tr>
</tbody>
</table>

The VBGM parameters, estimated by a least square procedure, are summarized in Figure 8. Longevity estimated by using inverted VBGM was 17.7 months and is in agreement with modal progression analysis. Asymptotic length (L∞= 29.6mm) estimated was close to largest size observed in the wild (30.49mm) (Figure 11).

Reproductive females observed during summer were likely hatched in previous spring and ranged from 4 to 6 months old. Autumn also presented a limited number of adult age classes (4 to 6 months old) as well as reduced frequency of ripe females and CPUE of adult females.

No recruitment peak was recorded during early winter months, despite of the presence of large (4 to 10 months old) and ripe females in the area. During late winter (September), a marked increase in recruitment density was observed, suggesting that spawning season starts during previous month with younger females.
However, spawning reaches a maximum intensity in October when cohort hatched in previous spring reaches one year old. Late winter spawning is also confirmed by the presence of very small recruits in October, suggesting spawning activity during the previous month. Nevertheless, spring presented highest relative abundance of adult females, percentage of ovary maturation as well as larger females contributing to recruitment (4 to 12 months) (Figure 10). The importance of recruitment originated from October spawning is confirmed by high abundance of recruits during the following month (November), especially in class interval of 7 mm (CL) or the equivalent age of approximately 1 month (Figure 10).

In October, the highest CL mean value was verified, even with significant frequency of juveniles in length classes between 4 and 15 millimeters. The smaller (CL= 4 mm) were probably hatched in October and the largest (CL= 15 mm) in August. A marked decrease in mean CL was observed in November. This reduction in mean CL is explained by high frequency of small individuals (peak at 7 mm), originated from October spawning (Figure 10).

TABLE III – Statistic summary of *A. longinaris* stock-recruitment data fitted to the models, containing the estimated parameters with standard deviations (within brackets) and coefficient of determination ($R^2$). Significant adjusts are indicated by * and non significant by ns. Overall p-values (p(OV)) are also provided. B&H= Beverton and Holt model, OV= overall, GE= good environment and BE= bad environment.

<table>
<thead>
<tr>
<th>Model</th>
<th>a</th>
<th>b</th>
<th>K</th>
<th>$R^*$</th>
<th>p (OV)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Ricker (OV)</td>
<td>16.1(6.37)</td>
<td>0.01 (0.004)</td>
<td>-</td>
<td>0.44</td>
<td>0.03</td>
</tr>
<tr>
<td>Ricker (GE)</td>
<td>19.9 (8.13)$^{ns}$</td>
<td>0.01 (0.004)$^{ns}$</td>
<td>-</td>
<td>0.54</td>
<td>0.15</td>
</tr>
<tr>
<td>Ricker (BE)</td>
<td>10.4 (2.31)</td>
<td>0.01 (0.003)</td>
<td>-</td>
<td>0.95</td>
<td>0.01</td>
</tr>
<tr>
<td>B &amp; H (OV)</td>
<td>0.026 (0.02)$^{ns}$</td>
<td>0.002 (0.0006)$^{ns}$</td>
<td>-</td>
<td>0.48</td>
<td>0.02</td>
</tr>
<tr>
<td>B &amp; H (GE)</td>
<td>0.005 (0.002)$^{ns}$</td>
<td>0.002(0.00008)$^{ns}$</td>
<td>-</td>
<td>0.98</td>
<td>0.001</td>
</tr>
<tr>
<td>B &amp; H (BE)</td>
<td>0.07(0.03)</td>
<td>0.003(0.0005)</td>
<td>-</td>
<td>0.91</td>
<td>0.002</td>
</tr>
<tr>
<td>Shepherd (OV)</td>
<td>7.99(16.9)$^{ns}$</td>
<td>0.97(5.7)$^{ns}$</td>
<td>74.3(933)$^{ns}$</td>
<td>0.20</td>
<td>0.45</td>
</tr>
<tr>
<td>Shepherd (GE)</td>
<td>19.99(40.1)$^{ns}$</td>
<td>0.93(3.0)$^{ns}$</td>
<td>33.2(353)$^{ns}$</td>
<td>0.30</td>
<td>0.41</td>
</tr>
<tr>
<td>Shepherd (BE)</td>
<td>9.0(4.6)</td>
<td>1.49(0.24)$^{ns}$</td>
<td>74.6 (29.8)$^{ns}$</td>
<td>0.95</td>
<td>0.01</td>
</tr>
</tbody>
</table>

Results obtained from overall multiple linear regression suggested that main factors influencing recruitment of *A. longinaris* in southern Brazil are salinity in spring ($\beta = 0.96; p= 0.05$) and rainfall during recruitment ($\beta = -3.29; p= 0.05$), which indirectly results in lower salinity during recruitment. When the favorable environment years were analyzed separately, main factor affecting recruitment was salinity during spawning season ($\beta = 0.99; p= 0.00$). During adverse environment years the main factor regulating recruitment biomass was the abundance of spawning stock ($\beta = 1.01; p= 0.03$). Total amount of rain in spawning year was not included in the model, since no significant p-value was obtained. Table IV
TABLE IV – Summary of multiple regression analysis estimated for *A. longinaris* in Southern Brazil. Three different models were obtained and significant p-values included are highlighted. SALS = salinity during spring (spawning season), RRC = rain during recruitment to commercial fishery, TRAIN = total rain during the spawning year and SS = spawning stock biomass.

<table>
<thead>
<tr>
<th></th>
<th>Beta</th>
<th>Std. Err.</th>
<th>B</th>
<th>Std. Err.</th>
<th>t</th>
<th>p</th>
<th>R² adj</th>
</tr>
</thead>
<tbody>
<tr>
<td>SALS</td>
<td>0.96</td>
<td>0.41</td>
<td>12.29</td>
<td>5.23</td>
<td>2.35</td>
<td>0.05</td>
<td>0.90</td>
</tr>
<tr>
<td>RRC</td>
<td>-3.29</td>
<td>1.41</td>
<td>-1.25</td>
<td>0.53</td>
<td>-2.34</td>
<td>0.05</td>
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</tr>
<tr>
<td>TRAIN</td>
<td>3.15</td>
<td>1.69</td>
<td>0.75</td>
<td>0.40</td>
<td>1.87</td>
<td>0.10</td>
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</tr>
</tbody>
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**bad years**

<p>| | | | | | | | |</p>
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<tr>
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</thead>
<tbody>
<tr>
<td>SS</td>
<td>1.01</td>
<td>0.18</td>
<td>3.13</td>
<td>0.55</td>
<td>5.72</td>
<td>0.03</td>
<td>0.88</td>
</tr>
<tr>
<td>TRAIN</td>
<td>0.24</td>
<td>0.18</td>
<td>0.12</td>
<td>0.08</td>
<td>1.39</td>
<td>0.30</td>
<td></td>
</tr>
</tbody>
</table>

**good years**

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<thead>
<tr>
<th></th>
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<th></th>
<th></th>
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</tr>
</thead>
<tbody>
<tr>
<td>SALS</td>
<td>0.99</td>
<td>0.10</td>
<td>15.92</td>
<td>1.54</td>
<td>10.32</td>
<td>0.00</td>
<td>0.88</td>
</tr>
</tbody>
</table>

DISCUSSION

Seasonal reproductive dynamics and size at first maturity

*Artemesia longinaris* clearly shows a year round spawning pattern, with main peak taking place in early spring (October). Preparation for main spawning event starts in winter, when increasing frequency of developing and ripe females were observed in surrounding area of Patos Lagoon Estuary. Furthermore, larger females, hatched during previous spring (12 months) participate in this reproductive event, which also increases the chances of successful recruitment (Vance *et al*. 1998).

Female penaeid shrimps usually present strong relationship between fecundity and size (Penn 1980). For instance, reared individuals of *A. longinaris* weighing 5g produce more than 78000 eggs. Conversely, a female weighing 2g produces only 1150 eggs (Scelzo 1991). Therefore, the presence of larger, as well as greater abundance of ripe females during October, explains the higher success of cohort hatched during spring.

Other two reproductive events, in autumn and in summer, are less important for population renewal and they seem to be more variable in intensity and period of occurrence (Ruffino & Castello 1992, Calazans 2002). Calazans (2002) reported autumn and spring as the seasons that concentrate similar high densities of *A. longinaris* post-larvae. Conversely, during the studied year, a much lower frequency and abundance of large and ripe females was recorded during autumn. Smaller females, hatched in summer and spring participated on this event, which was evidenced by size frequency and growth analysis. As stated by several authors (Guilland & Rothschild 1981, Vance *et al*. 1998) penaeid shrimps inhabiting tropical and sub-tropical regions usually present two recruitment peaks, one in spring and a secondary in autumn, always connected to phytoplankton blooms. However, a marked interannual variation in strength, area and time of secondary reproductive events is very often observed in penaeid populations and it is usually explained by unfavorable environmental conditions (Vance *et al*. 1998).

It is important to stress out that during autumn/2002 and winter/2002, the Southern Brazilian coast was under the influence of El Niño Southern Oscillation (ENSO) (Wolten & Timlin 1993, 1998), increasing rainfall and consequently reducing salt concentration in surrounding area of Patos Lagoon Estuary (Garcia *et al*. 1996). Dumont (2005), observed that *A. longinaris* tends to avoid lower salinity areas, under the influence of freshwater outflow from Patos Lagoon, which may have caused recruitment failure during this period.

The summer reproductive event is recognized as less important for population replenishing, since very low densities of post-larvae (Calazans 2002), adult and ripe females are reported. Females participating in this event are younger and originated from spring spawning (4 to 6 months) reaching final
maturation during summer. Higher water temperature seems to have a negative effect on summer reproduction. Growth of A. longinaris reared at 26 °C is very low and mortality rate is high (89%), suggesting that this temperature is close to lethal for this species (Haran et al. 1992). In summer, mean water temperature reached 25.64 °C, approaching to lethal value reported for this species (Haran et al. 1992). Therefore, reproduction of A. longinaris in the coast of Rio Grande do Sul seems to be constrained during summer, since under relatively high temperatures larval production and survival of cohort hatched in this period is reduced (Haran et al. 1992).

Previous investigation performed in Argentinean coast concluded that reproductive activity is closely related to water temperature (Petriella & Bridi 1992, Castillo et al. 2007). These investigations stress that ovary development is interrupted in autumn, synchronized with decreasing temperature. Spawning in Argentina is observed again only in December, when water temperature reaches 17°C (Petriella & Bridi 1992). Therefore, population of A. longinaris inhabiting Southern Brazil presents an extended reproductive season, which is linked to the latitude and consequently the temperature regime of the region. Several penaeid shrimp species distributed in higher latitudes tend to present well defined reproductive seasons, constrained to few months (Dall et al. 1990), while those species that live in warmer latitudes tend to present ripe ovaries and spawning activity all year round (Gulland & Rothschild 1981, Buckworth 1985, Crocos 1987). Briefly, reproductive cycle of A. longinaris coincides with pattern suggested to other penaeid shrimps, presenting a more stable spawning season in spring and secondary events more susceptible to environmental variation (Vance et al. 1998).

Besides the influence on the period of the reproductive cycle, temperature also plays an important role on size at first maturity, in such a way that populations distributed in higher latitudes have later gonadal development and therefore reach larger LM (Courtney & Massel 1997). Comparison of mean length at first maturity of population inhabiting Southern Brazil (16.76mm) to investigations performed in northern and southern limits of distribution confirms this pattern. The LM estimated for population inhabiting 23°S was 13.6mm, while the estimate for southern limit (37°59’S) was 22.1mm (Castillo et al. 2007). In fact, shrimp populations inhabiting higher latitudes tend to present slower growth rates and consequently delaying size at first maturity (Gulland & Rothschild 1981).

Differential length composition by depth, suggests a reproductive migration of this species to deeper waters. Thus, it is likely that females move away from shallow waters to deeper fishing grounds (16-20m) to complete ovary maturation (Boschi 1969), avoiding areas under the influence of estuarine outflow (Dumont 2005). Spawning seems to occur in similar depths, since highest larval abundance was observed between 10 and 30 isobaths, decreasing from 30 to 60m (Calazans 2002).

Migration of adult females to deeper waters was also observed in Argentina. During main reproductive season (summer) no mature females are found in artisanal fishing grounds (0-10 meters), suggesting a reproductive movement to deeper waters (Boschi 1969, Castillo et al. 2007). This pattern of reproductive migrations have been verified for several penaeid shrimps (Dall et al. 1990).

**Stock-recruitment and multiple regression analysis**

Stock-recruitment analysis (SRR) showed considerable interannual variations, which are not connected to any obvious trends in spawning stock biomass. It is quite clear that whatever the relationship between spawning stock and recruitment biomass, it is also influenced by environmental factors. Therefore, it is convenient to describe stock-recruitment relationship by a family of curves, each corresponding to an environment condition (Gulland & Rothschild 1981). In the present paper we suggest the use of two different groups of SRR, one for favorable environmental conditions (good year), with higher recruit biomass ratio and another curve for lower recruitment ratio (bad year). Visual classification of spawning/recruit data was supported by significant differences in salinity during spawning season, in such a way that good years presented higher salinity and reduced rainfall. Salinity has been reported as the main factor influencing the recruitment success of penaeid shrimps (Garcia 1981) and it is especially
true for penaeid species that cannot tolerate low salinities, where life cycle does not depend on an estuarine life stage (Anger 2003).

Changes in salinity can affect feeding activity as well as growth of crustaceans, altering metabolic pathways and developmental processes (Dall et al. 1990). This becomes especially conspicuous on yolk utilization during embryonic development, which may eventually cause significant intraspecific variation in the size and biomass of late eggs and early larvae (Giménez & Anger 2001). Such mechanisms should therefore, affect viability and, in particular, the nutritional vulnerability of planktotrophic larvae, thus playing a significant role for chance of larval survival in species living under variable salinity regimes (Giménez 2002, Giménez & Anger 2003).

Concern has been expressed that management decisions in the past had mostly been made on the basis that recruitment numbers are independent of parental stock abundance at levels of exploitations being experienced, in such a way that only environmental aspects regulate the recruitment success (Gulland & Rothschild 1981). However, this statement may not hold true for all shrimp fisheries (D’Incao et al. 2002), since high exploitation rates combined with unfavorable environmental conditions can cause dangerous reductions in spawning biomass. This is clearly true for A. longinaris, since highly significant fits to SRR models were observed under bad environmental conditions.

The Beverton & Holt model showed the best fit to spawning and recruitment biomass and as previously stated describes a relationship where an asymptotic density of spawning biomass is achieved, stabilized by density-dependent factors. The best fit of Beverton & Holt model to shrimp data was also observed in other shrimp fisheries worldwide (e.g. Cheng 1981), agreeing with results obtained in this investigation. Despite of problems reported for stock-recruitment relationships, this model can be used to detect recruitment overfishing, especially under very low spawning biomass levels, as well as to understand the resilience of studied population (Maury 1996).

The preliminary predictive model developed permits to conclude that salinity at spawning season and rain at recruitment explain 90% of recruitment success of A. longinaris in surrounding area of Patos Lagoon when all years are pooled together. Salinity is widely known as a limiting factor for penaeid shrimp reproduction and most of species depend on higher salinity to maturate the ovaries (Dall et al. 1990). The amount of rainfall during recruitment season is clearly linked to reductions in salinity of studied area (Garcia 1996), resulting in a negative effect on cohort survival. Since in this group, most metabolic and physiological processes (e.g. reproduction) require a stable osmotic and ionic environment, salinity is generally considered a key ecological factor in coastal and estuarine areas (Anger 2003).

The negative effect of brackish water on stenohaline coastal species, such as A. longinaris, is that under suboptimal salinity concentration the energetic requirements are increased due to hyper-osmoregulation (Kinne 1971). Allocation of the energy initially designated to reproduction is diverted to osmotic regulation and may inhibit ovary maturation of shrimps, since it is a high energy demanding process (Dall et al. 1990).

Suboptimal salinities also play a negative role on decapod crustacean larvae. Brackish water can slow down the molt interval and reduce growth, delaying larval development (Anger 2003). Even a slight delay in larval development may further reduce the chance of survival in plankton where slowly developing larvae are exposed for longer period to potentially harmful factors, such as physical stress, food limitation and pelagic predation (Lalli & Parsons 2006).

Results obtained from bad year regression reinforce those estimated through SRR. As previously mentioned, dangerous spawning biomass levels can be achieved under unfavorable environmental conditions, and preservation of biomass during these periods is vital to keep the stock healthy. Conversely, salinity is the main factor during those years considered as favorable. Even under low spawning biomass, the high fecundity (Scelzo 1991), growth rate (Ruffino & Castello 1992) and continuous spawning along the year allow recruitment success for this species.

Results obtained from multiple regressions must be carefully interpreted and only provide a preliminary idea about main factors affecting SRR, since a few number of seasons were analyzed. Therefore, further investigation must be developed using more variables and a greater number of years to identify new
environmental factors causing trends in stock biomass. Nevertheless, results obtained in this investigation confirm salinity as key factor for reproductive success of *A. longinaris* in surrounding area of Patos Lagoon.

Once reproductive seasons analyzed in present investigation showed marked differences in terms of importance to population replenishing, future analysis can be refined assuming heterogenic reproductive performances along the year and differential depth distribution according to size. The evident susceptibility to environmental factors and high natural mortality rates (Ruffino & Castello 1992) combined with excessive fishing effort may cause serious harm to the stock of *A. longinaris* in southern Brazil.

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